

AIRBORNE PARTICULATE MATTER – CHALLENGES FOR ABATEMENT

Roy M. Harrison
University of Birmingham
and

UK National Centre for Atmospheric Science

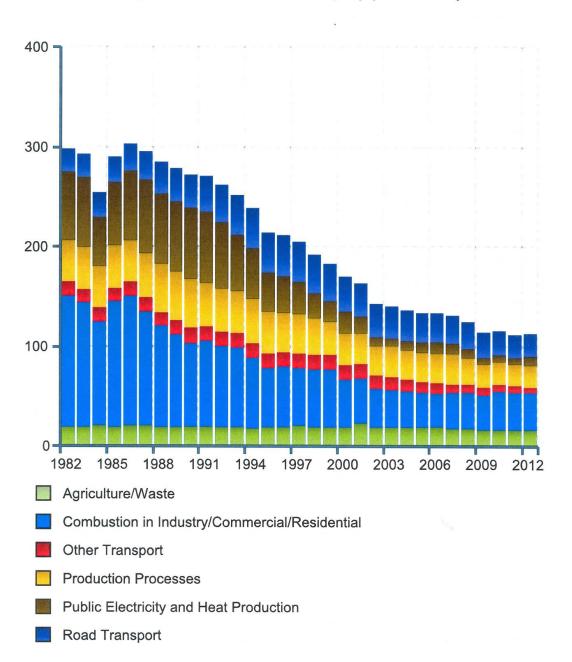


CONTENT

- Sources of particulate matter
- Receptor modelling of particulate matter
- Abatement issues
- Exposure reduction

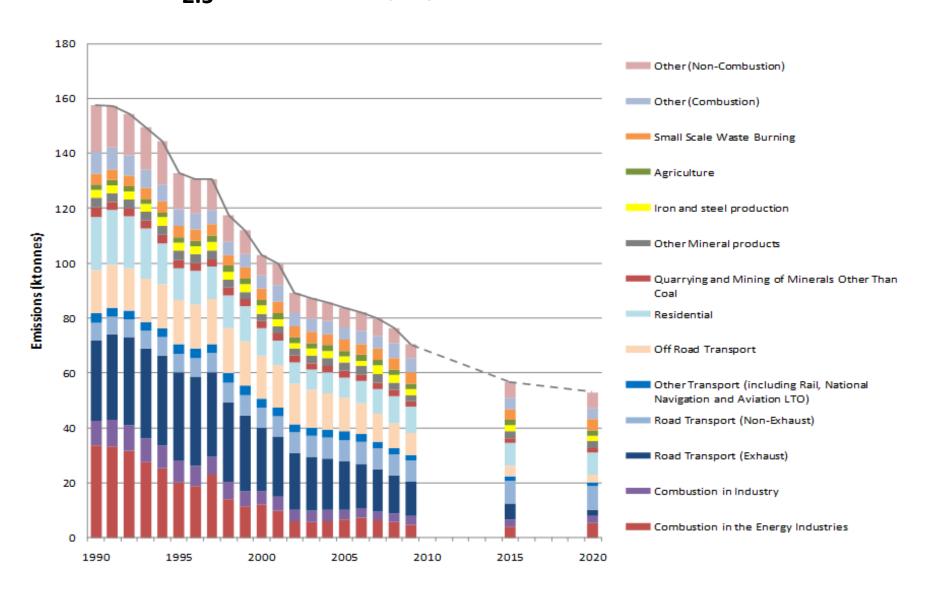


PM 10 (Particulate Matter < 10μm) (kilotonne)

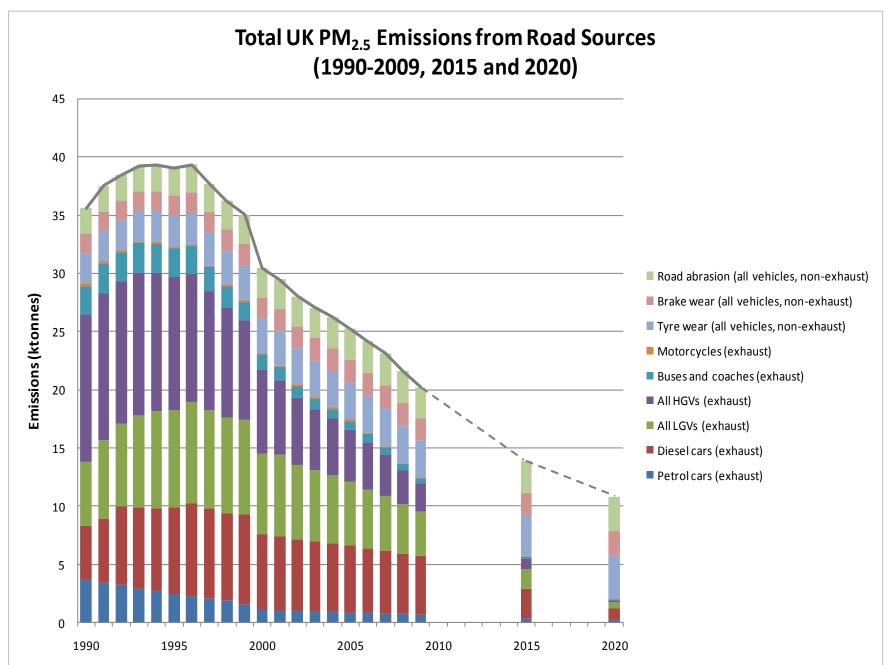




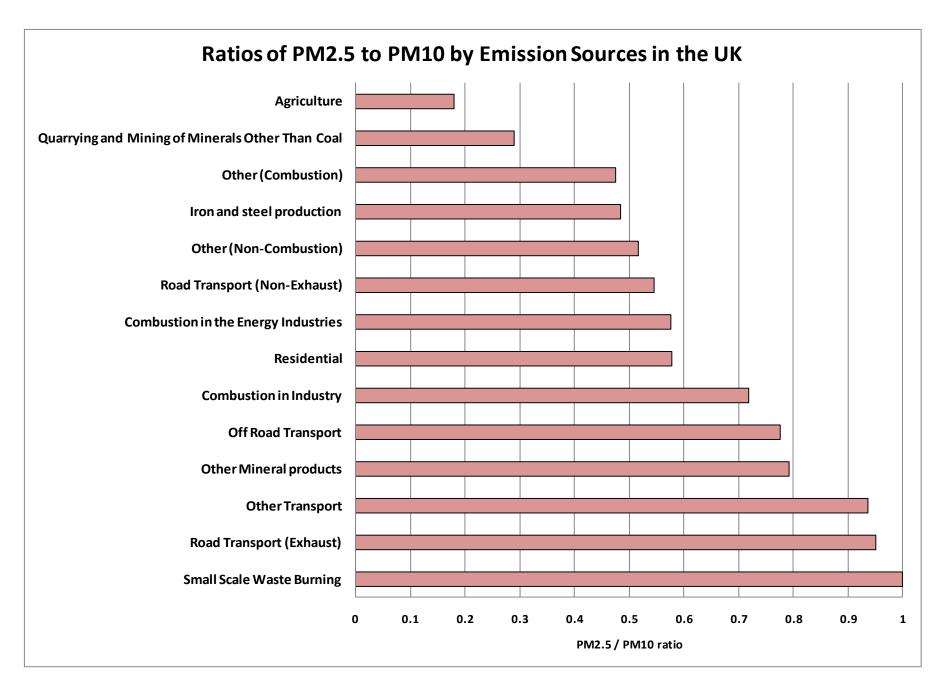
Total PM_{2.5} emissions (kt), 1990-2009, 2015 and 2020













RECEPTOR MODELLING

- Use of air quality data to infer the sources responsible for measured pollution levels (opposite of dispersion modelling!)
- Receptor modelling of airborne particles depends upon an assumption of mass conservation

$$C_i = \sum_{f_{i,j} g_j}^{j}$$

where $C_i = \text{airborne concentration of component, } i$ $f_{i,j} = \text{mass fraction of component } i \text{ in particles from source, } j$ $g_j = \text{mass of particles from source } j \text{ in an air sample}$

 Analysis of many air samples for multiple chemical components is necessary



TYPES OF RECEPTOR MODELLING OF PARTICULATE MATTER

There are two main types

Chemical Mass Balance

- Requires only one air sample, although better results are obtained with more
- Requires knowledge of chemical composition of particles from each source ($f_{i,j}$)
- Varies g_i for all chemical components to obtain best fit to mass conservation equation

Multivariate Statistical

- Principal Component Analysis widely used, but Positive Matrix
 Factorization (PMF) has advantages and is more frequently utilised
- Requires no advance knowledge of source chemical composition
- Requires many separate samples, and identifies temporal correlations of components (e.g. Na and Cl in sea salt) in a multidimensional space.



RECEPTOR MODELLING WITH CMB MODEL

 Uses organic molecular markers and trace elements to apportion the carbonaceous component of PM_{2.5}

 Source apportionment of the entire PM_{2.5} is conducted using the Pragmatic Mass Closure Model

 Results have been processed for winter air samples collected at LNK and HAR



CHEMICAL MASS BALANCE STUDY USING MOLECULAR MARKERS

- PM_{2.5} samples were collected and analysed for
 - \triangleright *n*-alkanes from $C_{24} C_{36}$
 - > 9 specific hopanes
 - > 13 PAH
 - > 14 carboxylic acids
 - levoglucosan
 - cholesterol
 - inorganic marker elements (Si, Al)

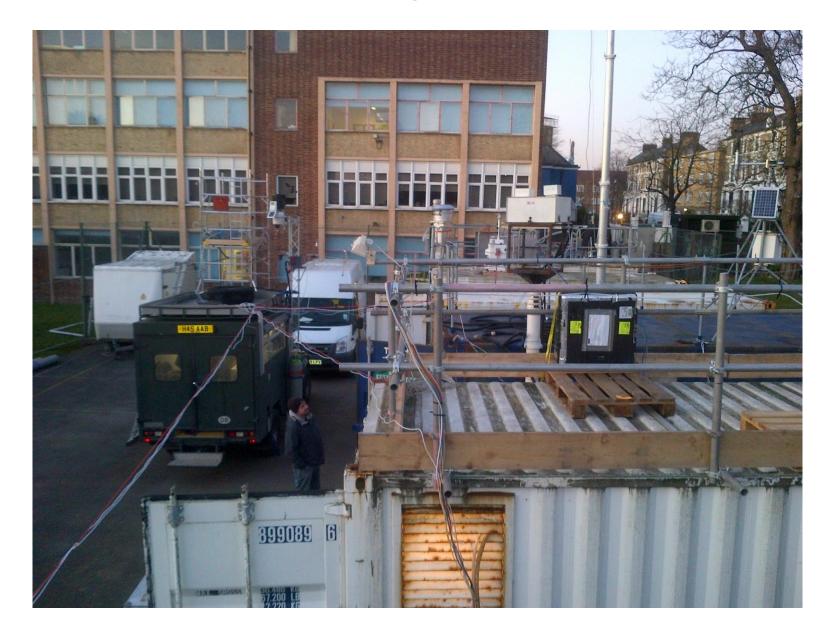


CMB MODEL RESULTS

- Model used to apportion sources of organic carbon to:
 - > road traffic
 - > vegetative detritus
 - dust and soil
 - > wood smoke
 - > coal combustion
 - natural gas combustion



NK Site During ClearfLo (1)



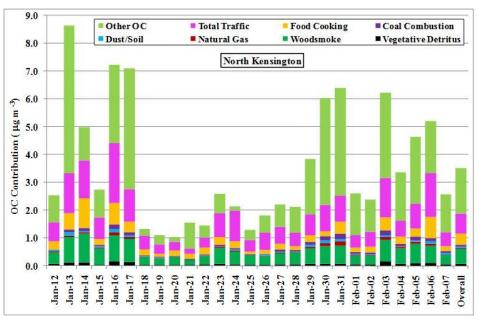


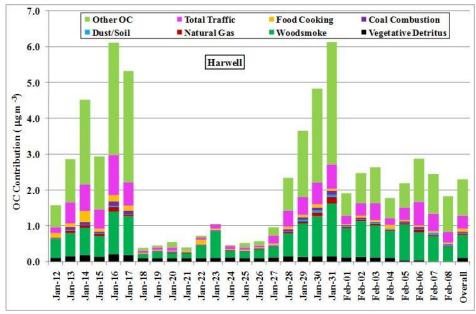
NK Site During ClearfLo (2)





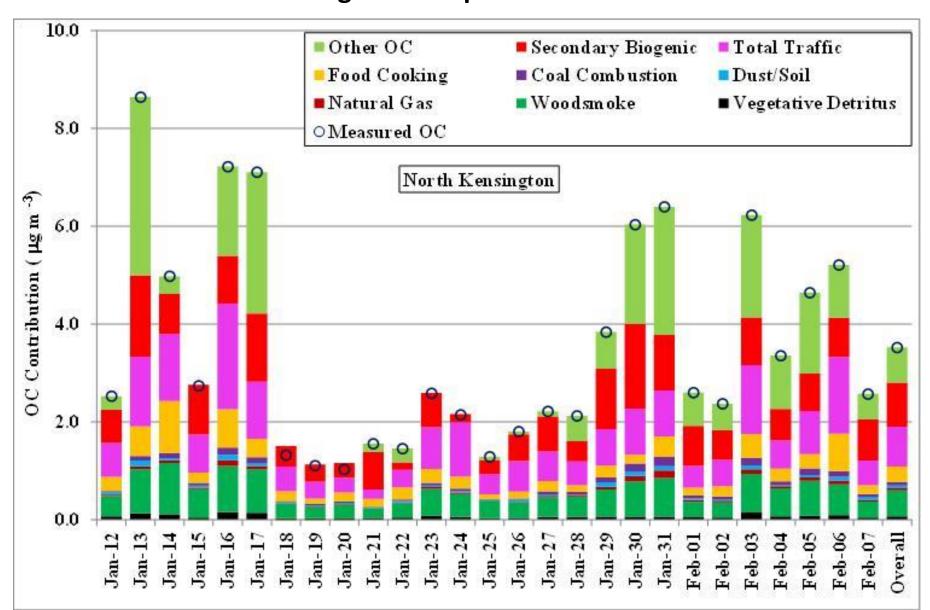
Daily OC Source Contributions at NK and HAR





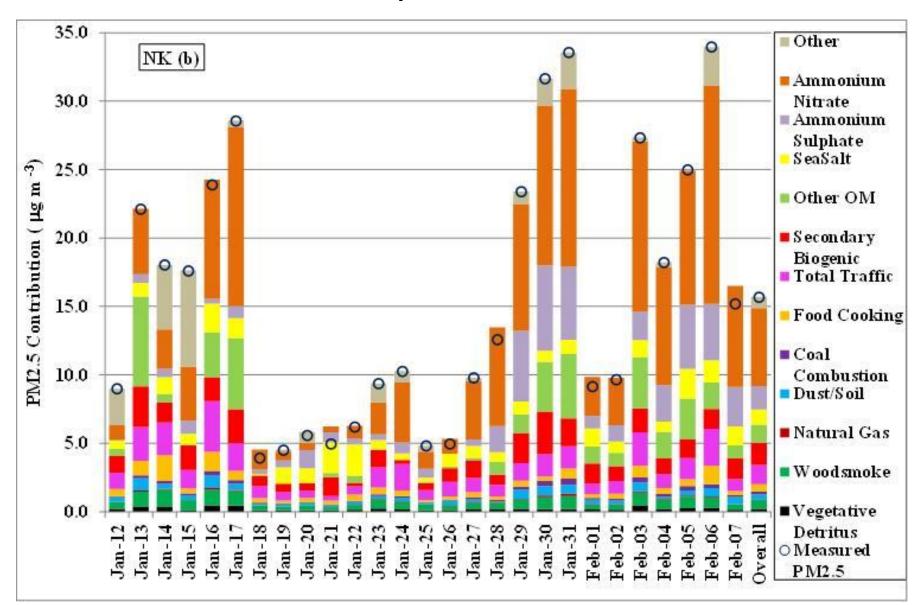


Daily OC Source Contribution Estimates with Secondary Biogenic Components at NK



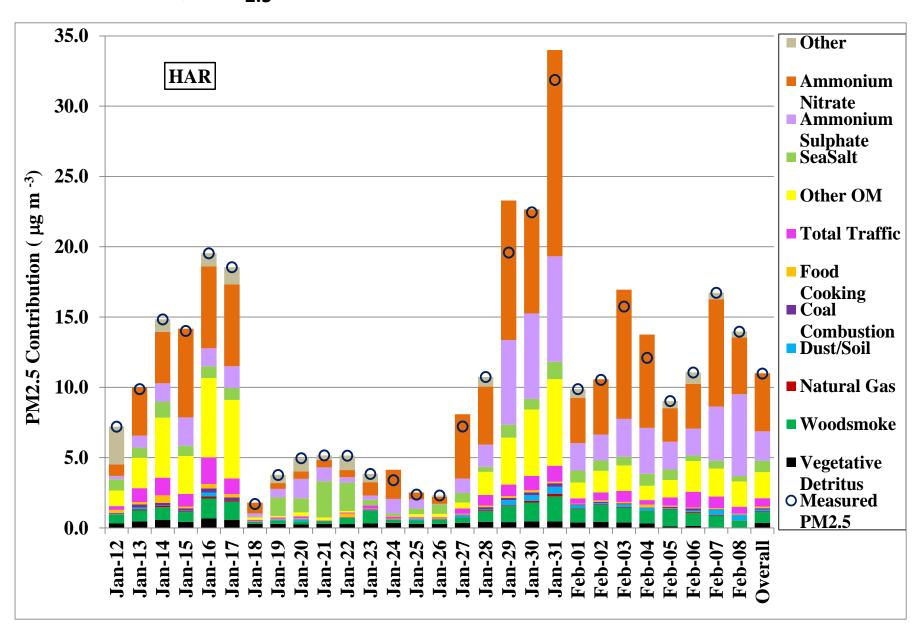


Daily PM_{2.5} Source Contribution Estimates with Secondary Biogenic Components at NK





Daily PM_{2.5} Source Contribution Estimates at HAR





SELECTED MEAN CONTRIBUTION TO PM_{2.5} MASS, μg m⁻³ (%)

	NORTH KENSINGTON	HARWELL
COOKING	0.69 (4%)	0.13 (1%)
WOODSMOKE	0.64 (4%)	0.76 (7%)
TRAFFIC EXHAUST	1.26 (8%)	0.61 (6%)
SULPHATES AND NITRATES	8.0 (51%)	6.2 (56%)
PM _{2.5} MASS	15.7	11.0



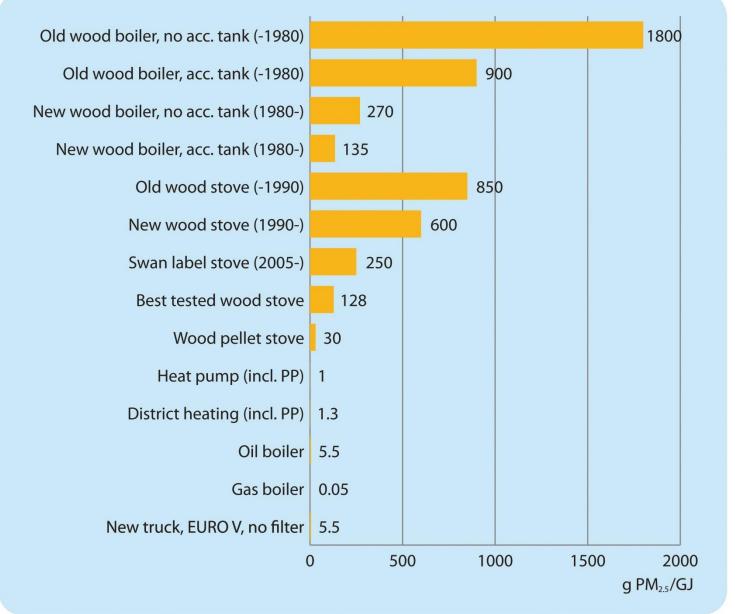
SPECIFIC SOURCES: WOODSMOKE

 Concentrations are relatively uniform across southern England.

 Although comprehensive evidence is lacking, it seems probable that emissions are increasing.

• Emissions inventories have a major problem in quantification of residential wood burning.

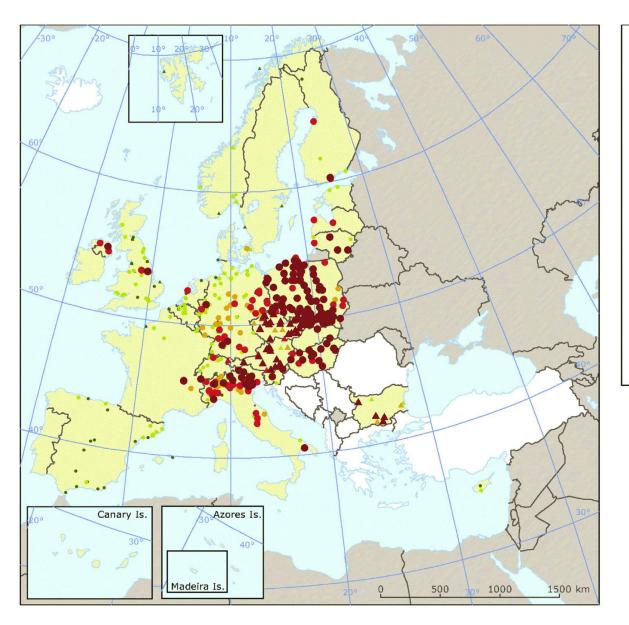




PM_{2.5} emission levels including condensates (Norwegian standard NS 3058-2) from boilers and stoves compared to other heat sources. In comparison, the emission level of a truck without filter (EURO V) is included.



Annual mean concentration of benzo(a)pyrene (ng/m³), 2011 (EEA, 2013a)



Annual mean benzo(a)pyrene 2011, based on annual average with percentage of valid measurements ≥ 14 % in ng/m³

- O Reported in PM10 fraction
- △ No indication of PM10 fraction
 - ≤ 0.12
 - 0.12-0.4
 - 0.4-0.6
 - 0.6-1.0
 - ▶ > 1.0
 - No data
 - Countries/regions not included in the data exchange process



AIR QUALITY POLLUTANTS

Polycyclic aromatic hydrocarbons

□ The EU air quality target (1 ng m⁻³ of benzo(a)pyrene) equates to a lifetime risk of 1 in 10⁴ which is 10 to 100-fold higher than that normally considered tolerable in regulatory toxicology



SPECIFIC SOURCES: COOKING AEROSOL

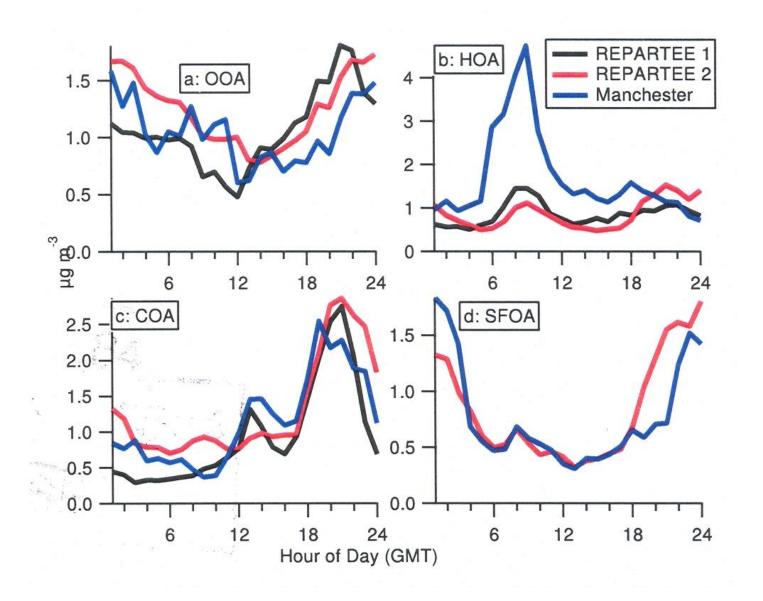
 Studies using Aerosol Mass Spectrometers (AMS) have highlighted this source.

 There are indications that the AMS data over-estimate the mass of cooking aerosol.

 Nonetheless, this is a source which cannot be ignored.



Median Diurnal Profiles of the Factors from the Three Campaigns (from J.D. Allan et al., ACP, 10, 647-668, 2010)





SPECIFIC SOURCES: NON-EXHAUST EMISSIONS FROM ROAD TRAFFIC

- Emission inventories include tyre wear, brake wear and road surface wear. They do not include particle resuspension.
- Currently, non-exhaust emissions of PM₁₀ are of a similar magnitude to exhaust emissions. By 2020, non-exhaust emissions will be strongly dominant.
- This source contributes similar masses of particles to the fine $(PM_{2.5})$ and coarse $(PM_{2.5-10})$ fractions.
- There are no current measures in place, or planned, to control emissions from this source.



ELEMENTAL DATA AS TRACERS OF NON-EXHAUST EMISSIONS

Examine:

- Relationship between metals to identify those with a common source
- Consider typical chemical origins of metals
- Fe, Cu, Sb and Ba characteristic of brake dust
- Al, Si, Ca, Ti are typically crustal and likely to arise from soil or resuspension
- Size distributions are indicative of source



MASS RECONSTRUCTION

Assumes

Brake dust = $Ba \times 91$

Tyre dust = $Zn \times 50$

Resuspension = $Si \times 3.6$

Gives contributions to mass of 0.9 – 11.5 μm particles

Brake dust = $55.3 \pm 7.9\%$

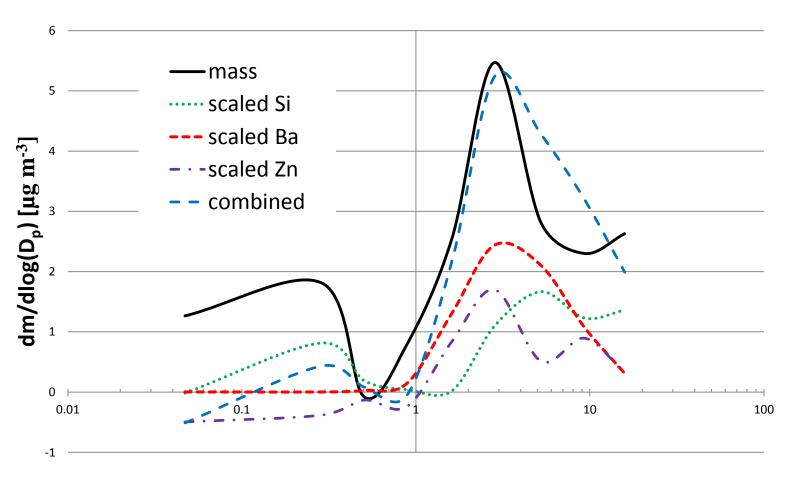
Tyre dust = $10.7 \pm 2.3\%$

Resuspension = $38.1 \pm 7\%$

Estimation of the Contribution of Brake Dust, Tire Wear and Resuspension to Nonexhaust Traffic Particles Derived from Atmospheric Measurements, R.M. Harrison, A. Jones, J. Gietl, J. Yin and D. Green, Environ. Sci. Technol., 46, 6523-6529 (2012).



Mass reconstruction - difference (all available data)



Diameter [µm]



SPECIFIC SOURCES: SECONDARY PARTICLES

- Sulphates and nitrates arise from the oxidation of sulphur dioxide and nitrogen dioxide respectively, and the precursorsecondary pollutant relationships appear to be strongly nonlinear.
- Abatement of ammonia (largely arising from agriculture)
 would be an effective way of slowing the oxidation of sulphur
 dioxide and hence reducing the formation of sulphate
 particles. It would also reduce nitrate particle formation.
- Most secondary organic aerosol derives from biogenic precursors, and is therefore unlikely to be subject to abatement measures.



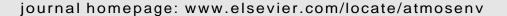


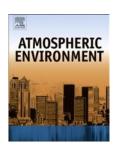
Atmospheric Environment 69 (2013) 211e218



Contents lists available at SciVerse ScienceDirect

Atmospheric Environment





The effect of varying primary emissions on the concentrations of inorganic aerosols predicted by the enhanced UK Photochemical Trajectory Model

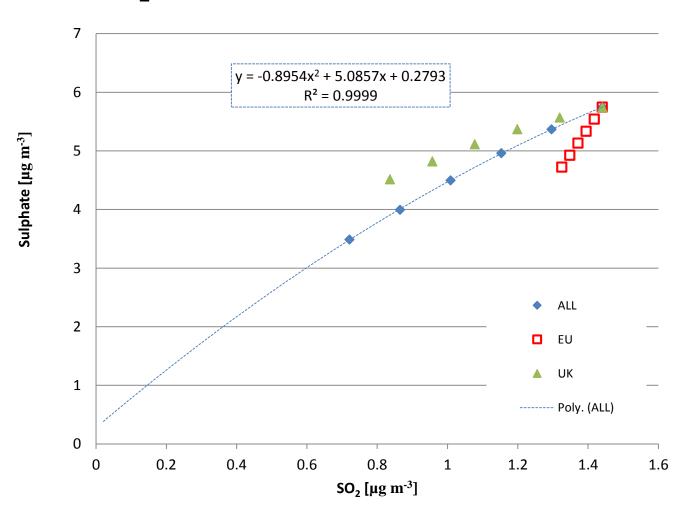
Roy M. Harrison*, Alan M. Jones, David C.S. Beddows, Richard G. Derwent 1

National Centre for Atmospheric Science, Division of Environment Health & Risk Management, School of Geography, Earth & Environmental Sciences, University of Birmingham, Edgbaston, Birmingham B15 2TT, United Kingdom





Predicted sulphate as a function of SO₂ at Harwell (SO₂ emissions varied in UK and Europe)

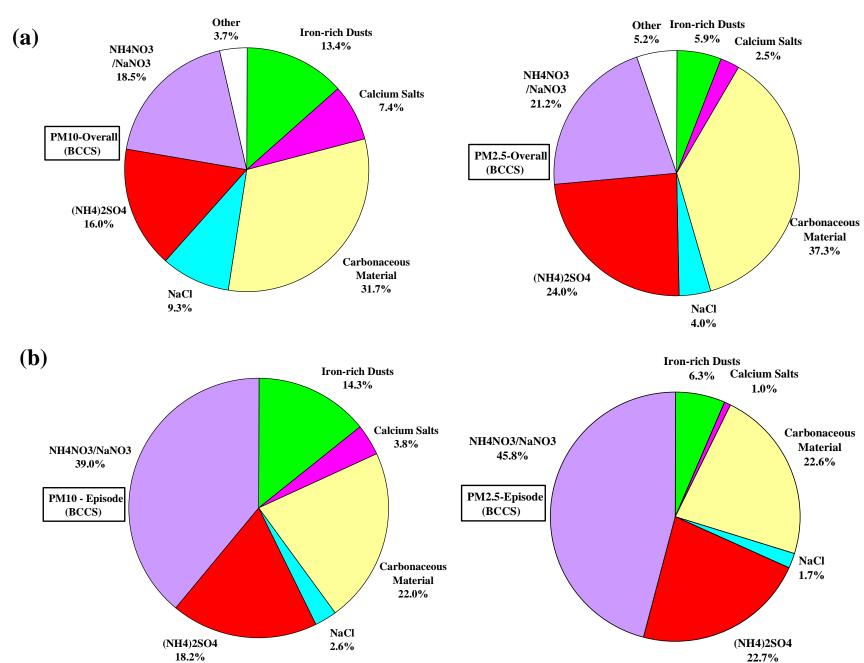




Days with Concentrations of $PM_{10} > 50 \mu g m^{-3}$

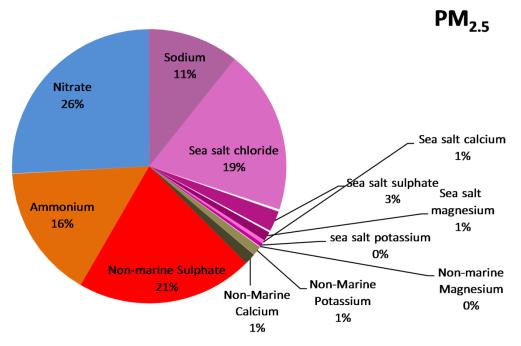
 As in earlier work, the component showing the greatest enhancement in concentration on high pollution days is nitrate in both PM₁₀ and PM_{2.5}

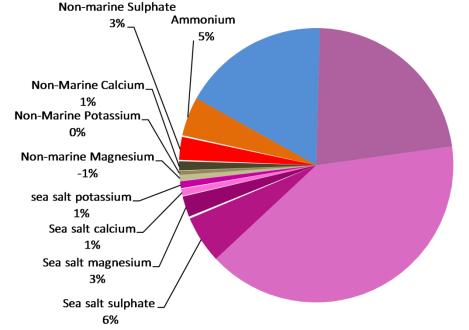






AUCHENCORTH MOSS, 2007-2012



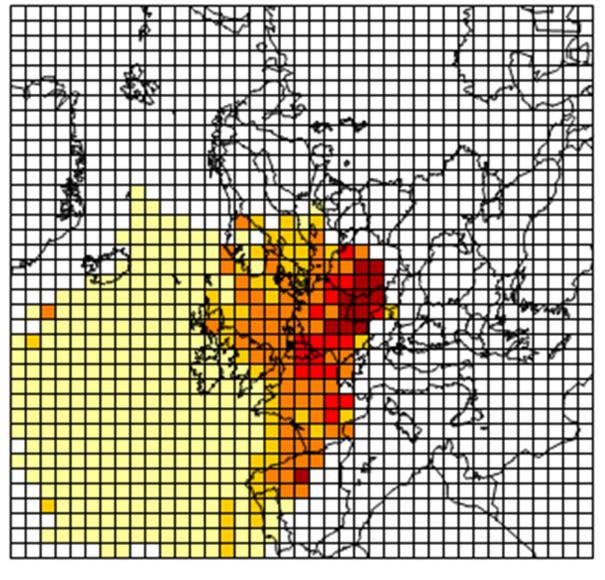


$\mathbf{PM}_{\mathrm{coarse}}$

Average composition by mass of the water soluble inorganic aerosol fraction measured by the MARGA from January 2007 to December 2012 in both the $PM_{2.5}$ and PM_{coarse} . (from Twigg et al., ACPD, 3703-3743, 2015).



NITRATE AT HARWELL

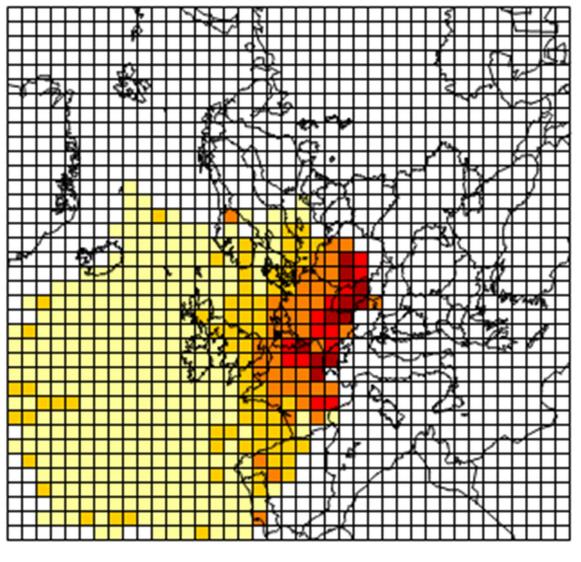


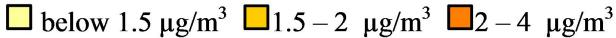
 \square below 1.5 μg/m³ \square 1.5 – 2 μg/m³ \square 2 – 4 μg/m³

 $\blacksquare 4 - 5 \text{ } \mu\text{g/m}^3 \blacksquare \text{ above } 5 \text{ } \mu\text{g/m}^3$



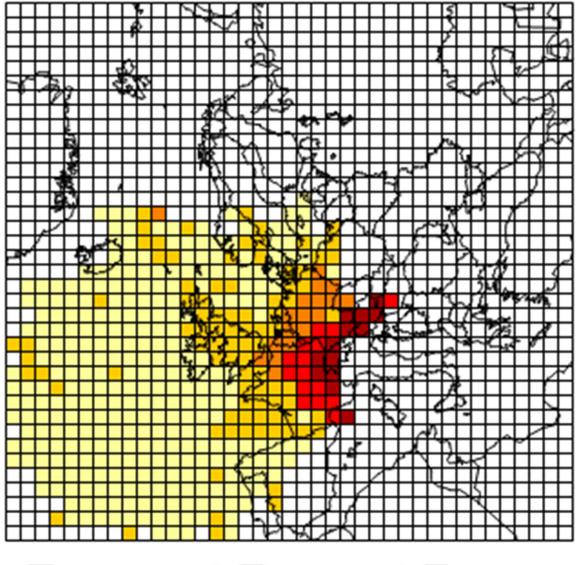
SULPHATE AT HARWELL





 $\blacksquare 4 - 5 \text{ } \mu\text{g/m}^3 \blacksquare \text{ above } 5 \text{ } \mu\text{g/m}^3$

PARTICULATE ORGANIC CARBON AT HARWELL



□below 2 μ g/m³ □2 − 3 μ g/m³ □3 − 4 μ g/m³

 $\blacksquare 4 - 5 \quad \mu g/m^3 \quad \blacksquare \text{ above 5 } \mu g/m^3$



EXPOSURE REDUCTION

- The EU approach to policy on PM_{2.5} has adopted the exposure reduction concept alongside traditional Limit Values.
- Exposure reduction implies maximising the function:

$$\sum^{i} \Delta P M_{i} \cdot N_{i}$$

- where Δ PM_i is the reduction in PM concentration in grid square i, and N_i is the population of grid square i.
- This implies that the most cost-effective policies may be those that focus upon reduction of urban concentrations, because of the large values of N_i in cities.



CONCLUSIONS

- Knowledge is improving of the sources contributing to $PM_{2.5}$ and PM_{10} in UK air.
- Sources such as wood smoke and cooking aerosol are significant and may lead to problems with abatement policies.
- Secondary particles are dominant and provide especially difficult mitigation issues due to non-linearity (sulphate and nitrate) and biogenic sources (organic aerosol).
- There are few local policy levers which reduce PM concentrations appreciably.

